# Phylogenetic relationships of *Atherina hepsetus* and *Atherina boyeri* (Pisces: Atherinidae) populations from Greece, based on mtDNA sequences

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The genetic divergence and the phylogenetic relationships of six *Atherina boyeri* (freshwater and marine origin) and five *Atherina hepsetus* populations from Greece were investigated using partial sequence analysis of 12s rRNA, 16s rRNA and control region mtDNA segments. Three different well divergent groups were revealed; the first one includes *A. boyeri* populations living in the sea, the second includes *A. boyeri* populations living in the lakes and lagoons whereas the third one includes all *A. hepsetus* populations. Fifty-seven different haplotypes were detected among the populations studied. In all three mtDNA segments examined, sequence analysis revealed the existence of fixed haplotypic differences discriminating *A. boyeri* populations inhabiting the lagoon and the lakes from both the coastal *A. boyeri* and the *A. hepsetus* populations. The genetic divergence values estimated between coastal (marine) *A. boyeri* populations and those living in the lagoon and the lakes are of the same order of magnitude as those observed among coastal *A. boyeri* and *A. hepsetus* populations. The results obtained by different phylogenetic methods were identical. The deep sequence divergence with the fixed different haplotypes observed suggests the occurrence of a cryptic or sibling species within *A. boyeri* complex. © 2007 The Linnean Society of London, *Biological Journal of the Linnean Society*, 2007, 92, 151–161.

ADDITIONAL KEYWORDS: 12srRNA – 16srRNA – D-loop – genetic divergence – lakes – lagoons – marine – sibling species.

# INTRODUCTION

The Atherinidae family is represented by a single genus in Europe consisting of three species, *Atherina boyeri* Risso 1810, *Atherina hepsetus* L. 1758 and *Atherina presbyter* Cuvier 1829 (Kiener & Spillman, 1969). *Atherina boyeri* is found over a wide range of temperature and salinity conditions, in coastal, estuarine, lagoon, and inland waters and it is subdivided to local semi-isolated populations that may be quite different in respect to their population biology and morphology (Henderson & Bamber, 1987).

Several studies based on morphometric data revealed differences among *A. boyeri* populations living in the sea compared to those living in the lagoons (Kartas & Trabelsi, 1990; Focant et al., 1992; Trabelsi et al., 2002a), whereas such differences were also existed between neighbouring marine populations (Trabelsi, Quignard & Kartas, 1994). A high degree of genetic differentiation has also been revealed through protein electrophoresis, among different A. boyeri marine populations (Berrebi & Britton-Davidian, 1980; Creech, 1991). Moreover, Focant, Rosecchi & Crivelli (1999) found different electrophoretic patterns sufficient to discriminate the populations living in lagoons from the marine ones. Consequently, A. boyeri is viewed by many as a polymorphic complex species. (Focant et al., 1999).

In our previous work (Klossa-Kilia *et al.*, 2002), high nucleotide divergence values have been detected, through restriction fragment length polymorphism (RFLP) analysis, among *A. boveri* populations living in

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the sea and those living in lagoons and lakes, leading to the conclusion that there are two different sibling species or at least subspecies or semispecies colonizing 'lagoon' and 'marine' habitats, respectively. The same conclusion was also reached by Trabelsi et al. (2002b) using a partial sequence of the mitochondrial cytochrome segment (cyt b) and they proposed to include the atherinids living in the lagoons (in France) in a separate new species (Atherina laguna), whereas they also distinguished the marine A. boyeri populations into two different species Atherina punctata and A. boyeri (depending on the presence or absent of dark spots along the lateral line, respectively). However, because the proposed taxonomic revision is based (at the molecular level) only on the sequence of a small mtDNA segment, and until this revision be widely accepted, we will still continue to use the current term for A. boyeri populations living in the lagoons or the sea. Recently Trabelsi et al. (2004) assumed the existence of different taxonomic units among A. boyeri populations living in different (and distant) lagoons. High genetic structure among A. boyeri populations originated from different lagoons of a wide geographical range has also been shown by Astolfi et al. (2005) and Congiu, Rossi & Colombo (2002), although the later authors found no significant differences among populations originated from different lagoons along the coasts of Adriatic Sea (possibly due to migration movements). The high degree of genetic differentiation among different A. boyeri populations revealed by all these studies has led to a confusion regarding the taxonomic status of this species, making it obvious that more data are needed to elucidate this problem. Moreover, the literature concerning phylogenetic relationships of A. boyeri with the other sympatric

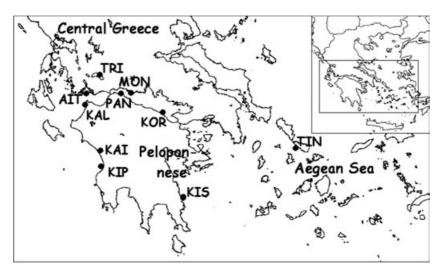
species, A. hepsetus, which prefer to live in the sea, is limited.

In the present study, we used the partial sequence of three mtDNA segments, a total of 1329 bp from 12s rRNA, 16s rRNA and control region, to investigate the phylogenetic relationships of A. hepsetus and A. boyeri populations from Greece. The majority of populations analysed were collected from different marine sites, while two populations of A. boyeri originated from different freshwater lakes and one from a lagoon. With this experimental approach, we aimed to: (1) explore the phylogenetic relationship among A. hepsetus and A. boyeri (lagoon, lakes and marine) Greek populations; (2) estimate the magnitude of divergence existing between A. hepsetus and A. boyeri populations in an attempt to determine the taxonomic status of the 'lagoon' A. boyeri populations; (3) ensure the high degree of nucleotide divergence exhibited among A. boyeri populations (marine versus lakes and lagoon), using sequencing analysis, in the light of its higher resolution power compared with RFLP analysis; and (4) determine whether there is any important genetic difference between A. boyeri populations living in distant isolated lakes.

# MATERIAL AND METHODS

#### SAMPLING

Five *A. hepsetus* populations and six *A. boyeri* ones, were collected from regions located mainly in western Greece (Fig. 1) using traps. Three *A. hepsetus* individuals were examined from each population originating from the coastal areas of Tinos (TIN), Panagopoula (PAN), Monastiraki (MON), and Kiparisi (KIS),



**Figure 1.** The sampling localities. AIT, Aitoliko lagoon; KAI, Lake Kaiafas; KAL, Kalamaki; KIP, Kiparissia; KIS, Kiparisi; KOR, Korinthiakos; MON, Monastiraki; PAN, Panagopoula; TIN, Tinos; TRI, Lake Trichonida.

**Table 1.** DNA sequence variability in the 12s rRNA region examined among the eight haplotypes

	11111	112222222	222333
	22302667	8822235555	578111
	2934120579	4902810123	430269
Ah12s1	ACATGCTTAG	AGTAGAGACC	TAAGTA
Ah12s2		C	
Ah12s3		C	
<b>Ab12s1</b>	.AC.	.ACGAG.T	AGCACC
Ab12s2	.AC.	.ACGAGTT	AGCACC
Ab12s3	.T.C.TCCGA	.ACG.G.GTT	AC.AC.
Ab12s4	GT.CATCCGA	.ACG.G.GTT	AC.AC.
Ab12s5	.TTC.TCCGA	.ACG.G.GTT	AC.AC.

Italics indicate lagoon type haplotypes.

whereas four were examined from Korinthiakos (KOR). Atherina boyeri specimens were sampled from the coasts of Kalamaki (KAL), Kiparissia (KIP) and Korinthiakos (KOR), the lakes Trichonida (TRI) and Kaiafas (KAI), and the Aitoliko lagoon (AIT) (Table 1). From each A. boyeri population, three individuals were used, except from the populations of Korinthiakos and Aitoliko, where five and four specimens, respectively, were used. All individuals were transfererred in dry ice to the laboratory and stored at -75 °C until use.

# DNA EXTRACTION, POLYMERASE CHAIN REACTION (PCR) AMPLIFICATION, AND DNA SEQUENCING

Total DNA was isolated from the muscle tissue according to the protocol included in the DNAeasy Tissue kit (Qiagen) and the resulting DNA was examined via agarose electrophoresis.

For the amplification of the three mtDNA regions studied, three sets of primers were used. The universal primers sets 12SAL-12SBH and 16SARL-16SBRH described by Palumbi et al. (1991) were used for the 12s rRNA and 16s rRNA segments, respectively, whereas we used the primer set CRA-CRB described by Lee et al. (1995) for the control region. PCR amplification included one preliminary denaturation step at 94 °C for 4 min followed by 35 cycles comprising denaturation at 94 °C for 1 min, annealing at 57 °C (12s rRNA)/50 °C (16s rRNA)/ 51 °C (control region) for 1 min, and primer extention at 72 °C for 1.5 min, ending with a final step at 72 °C for 5 min. The PCR products from the three regions amplified were confirmed via agarose electrophoresis and purified using the QIAquick PCR Purification kit (Qiagen). The light strand sequencing for the three mtDNA regions was performed by MWG Biotech-Germany.

#### SEQUENCE DATA ANALYSIS

We performed the analysis using each dataset (mtDNA segment) seperately as well as the combination of them. The sequence alignment was performed by Clustal W included in BioEdit, version 5.0.9 (Hall, 1999). Unread nucleotides and gaps were treated as missing data and as the fifth character, respectively. For each dataset, the evolutionary distances (d) were calculated among the 11 populations examined considering the Kimura 2 parameter (Kimura, 1980) model using MEGA, version 3.1 (Kumar, Tamura & Nei, 2004).

The proper substitution models for each dataset and for the combination of them were selected by Modeltest (Posada & Crandall, 1998). Neighbour-joining (NJ) trees (Saitou & Nei, 1987) were generated using MEGA and PAUP\* v4.b10 (Swofford, 2002). Maximum likelihood (ML) trees were generated using PAUP\* and considering the substitution models selected by Modeltest. Maximum parsimony (MP) analysis was also performed using PAUP\*. For the heuristic search, the stepwise addition and the branch swapping options were selected. For each option, a 50% majority rule tree was generated. For all trees plotted, bootstrap values (Felsenstein, 1985) were based on 10 000 pseudoreplicates. The merging of individuals haplotypes to populations in the combined data set was made using McClade, version 4.03 (Maddison & Maddison, 2001). To check the phylogenetic signal of the data the g<sub>1</sub> method (Hillis & Huelsenbeck, 1992) was applied. From the random trees option of PAUP\*, 10 million trees were evaluated.

## RESULTS

After PCR amplification and sequencing, 351 bp for 12s rRNA, 497 bp for 16s rRNA, and 481 bp for the control region were obtained. For the 12s rRNA segment, 26 variable (21 informative) sites were found. The 16s rRNA segment exhibited 69 variable (57 informative) sites whereas the control region 121 (115 informative) sites, respectively (Tables 1, 2, 3). The A. hepsetus individuals were sequenced and resulted in three, ten, and ten different haplotypes for the 12s rRNA, 16s rRNA, and control region, respectively, whereas A. boyeri individuals resulted in five, ten, and 19 haplotypes for the same segments, respectively (Table 4). No haplotype was shared between A. hepsetus and A. boyeri individuals; furthermore, no common haplotype was found between the coastal individuals of A. boyeri and those from the lakes and the lagoon. GenBank accession numbers for the 57 haplotypes we obtained are AY749048-AY749104.

The evolutionary distances calculated for the 12s rRNA region revealed that the marine *A. boyeri* pop-

Table 2. DNA sequence variability in the 16s rRNA region examined among the 20 haplotypes; dashes indicate deletions

	1111	1111111111	1111111111	1112222222	222222222	222222233	333333334
	345881114	5566677777	888888999	9990000011	1124455677	7778999901	111234594
	9390386789	4602912368	2346789023	7890356901	3907906302	5898014880	258811448
Ah16s1	TACAACGAGT	AGAATCAACC	-TCCCGGGGT	GGAAAACCCC	TACTACCCCA	GTCACTGTAT	GGTAAAAGT
Ah16s2			C				
Ah16s3			TCGTA				
Ah16s4	A		TCGTA	• • • • • • • • • •			
Ah16s5	G.A		CCTCG				
Ah16s6	A		C				
Ah16s7			CCTCG			A	A.
Ah16s8	A					T	G
Ah16s9			C				
Ah16s10							
Ab16s1	C.TAAC.C	GA.GGT.CT.	T.AG	CAGGTA	C.TGG.TTTG	ACA.GCG.	C.CTG.C
Ab16s2	C.TAAC.C	GA.GGT.CT.	T.AG	CA.GGGTA	C.TGG.TTTG	ACA.GCG.	C.CTG.C
Ab16s3	C.TAAC.C	GAGGGT.CT.	T.AG	CATTTT	C.TGG.TTTG	A.A.GCGG	C.CTG.C
Ab16s4	CGGA.GA.	.A.GCTG	G	GC-GG	CA.TT	.C.T.C.CG.	C.CG.CG
Ab16s5	CGGA.GA.	.A.GCTG	CG	GC-GG	CA.TT	.C.T.C.CG.	C.CG.CG
Ab16s6	CGGA.GA.	.A.GCTG	CC.G	GC-GG	CA.TT	.C.T.C.CG.	C.CG.CG
Ab16s7	CG.GGA.GA.	.A.GCTG	CCG	GC-GG	CA.TT	.C.T.C.CG.	C.CG.CG
Ab16s8	CGGAA.	.A.GCTGT	CCG	GC-GG	CA.TT	.C.T.C.CG.	C.CG.CG
Ab16s9	CGGA.GA.	.A.GCTG	CG	GC-GG	CC.A.TT	.C.T.C.CG.	C.CG.CG
Ab16s10	CGGA.GA.	.A.GCTG	CCG	GC-GG	CA.TT	.C.T.C.CG.	CACG.CGA.
ADIUSIU	CGGA.GA.	.A.GCIG	CCG		CA.11	.0.1.0.00.	CACG.CGA.

Italics indicate lagoon type haplotypes.

ulations differed almost equally from both the lagoon A. boyeri and the A. hepsetus populations, whereas the lagoon A. boyeri populations exhibited the greatest distance from the A. hepsetus ones. More specifically, A. hepsetus populations differed by 4.11-4.30% from the coastal A. boyeri ones and 5.38–5.79% from the lake and lagoon populations. Atherina boyeri coastal populations differed by 3.70-4.11% from the lake and lagoon ones (Table 5). Analogous differences were also observed for the 16s rRNA as well as for the control region and the combined data set (Tables 6, 7, 8), with the control reagion revealing the most extreme values as expected. In every case, the distance values estimated among A. boyeri populations (coastal versus lake and lagoon) are of the same order of magnitude compared with those estimated among the other two well defined species (A. boyeri and A. hepsetus).

The Akaike Information Theoretical Criterion (AIC) performed by Modeltest for the 12s rRNA haplotypes selected the Kimura 2 parameter model, with equal base frequencies, Ti/Tv = 2.4513 and a proportion of invariable sites (I) equaling to 0.8126. The same model was selected for the 16s rRNA haplotypes, with equal base frequencies, Ti/Tv = 3.4317 and I = 0.7552. The test for the control region haplotypes selected the HKY model with unequal base frequencies, Ti/Tv = 1.8293,  $\alpha = 0.8486$  and I = 0.5752. For the com-

bined dataset, the AIC criterion selected the TrN model with unequal base frequencies,  $\alpha = 0.8681$  and I = 0.6812.

The sequence data of the four dataset studied revealed a significant phylogenetic signal. The values calculated by the  $g_1$  method was  $g_1 = -0.4095$  (P < 0.001) for 12s rRNA,  $g_1 = -0.5075$  (P < 0.001) for 16s rRNA,  $g_1 = -0.4233$  (P < 0.001) for the control region and  $g_1 = -0.4121$  (P < 0.001) for the combined data.

For each mtDNA region studied as well as the combined data set, the NJ trees plotted resulted in the same topology. According to this, A. boyeri haplotypes are clustered in two groups, one containing the haplotypes from the coastal areas of KAL, KOR, and KIP, and the second containing the haplotypes from the lakes of TRI, KAI, and the AIT lagoon. All the A. hepsetus haplotypes were clustered together (Fig. 2). The MP trees plotted for each mtDNA region as well as for the combined data set (Fig. 3) resulted in the same topology, very similar to the topologies of the NJ trees. The same topology was also obtained with the ML analysis. In conclusion, in all distance, parsimony, and likelihood trees plotted, the lake and lagoon haplotypes of A. boyeri were clustered together, apart from the coastal A. boyeri haplotypes whereas A. hepsetus haplotypes were clustered together in a separate clade

Table 3. DNA sequence variability in the control region examined among the 29 haplotypes; dashes indicate deletions

Italics indicate lagoon type haplotypes.

**Table 4.** Sampling size and haplotype distribution per locality for the *Atherina boyeri* and *Atherina hepsetus* populations examined

	12	2s				16	s									Co	ont	rol																
Sampling locality*	1	2	3	4	5	1	2	3	4	5	6	7	8	9	10	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	3 19
Atherina boyeri haj	olo	typ	es																															
Kalamaki (3)	2	1										2	1			1	1	1																
Korinthiakos (5)	5													3	2													1	1	1	1	1		
Kiparissia (3)	2	1							1	1	1												1	1	1									
Aitoliko (4)			4			4													1	1	1	1												
Trichonida (3)			3			2	1															1				1	1							
Kaiafas (3)			1	1	1	2		1														1											1	1
Atherina hepsetus	naj	olot	typ	es																														
Tinos (3)	3					2	1														3													
Panagopoula (3)	3							2	1							1	1	1																
Monastiraki (3)	2	1					2			1									2	1														
Kiparisi (3)	3										1	1	1								1	1	1											
Korinthiakos (4)	3		1								2			1	1						2			1	1									

<sup>\*</sup>Numbers in parentheses indicate the number of specimens examined from each population.

**Table 5.** Genetic distances (%) (below the diagonal) and standard errors (%) (above the diagonal) calculated for 12s rRNA between the populations examined, considering the Kimura 2 parameter substitution model

	AhTin	AhPan	AhMon	AhKis	AhKor	AbTri	AbKai	AbAit	AbKor	AbKal	AbKip
AhTin		0.00	0.09	0.00	0.09	1.26	1.28	1.26	1.04	1.05	1.05
AhPan	0.00		0.09	0.00	0.09	1.26	1.28	1.26	1.04	1.05	1.05
AhMon	0.10	0.10		0.09	0.13	1.27	1.29	1.27	1.04	1.05	1.05
AhKis	0.00	0.00	0.10		0.09	1.26	1.28	1.26	1.04	1.05	1.05
AhKor	0.10	0.10	0.19	0.10		1.27	1.29	1.27	1.04	1.05	1.05
AbTri	5.38	5.38	5.48	5.38	5.48		0.16	0.00	1.07	1.05	1.05
AbKai	5.69	5.69	5.79	5.69	5.79	0.29		0.16	1.09	1.07	1.07
AbAit	5.38	5.38	5.48	5.38	5.48	0.00	0.29		1.07	1.05	1.05
AbKor	4.11	4.11	4.21	4.11	4.21	3.81	4.11	3.81		0.09	0.09
AbKal	4.21	4.21	4.30	4.21	4.30	3.70	4.01	3.71	0.10		0.12
AbKip	4.21	4.21	4.30	4.21	4.30	3.70	4.01	3.71	0.10	0.13	

**Table 6.** Genetic distances (%) (below the diagonal) and standard errors (%) (above the diagonal) calculated for 16s rRNA between the populations examined, considering the Kimura 2 parameter substitution model

	AhTin	AhPan	AhMon	AhKis	AhKor	AbTri	AbKai	AbAit	AbKor	AbKal	AbKip
AhTin		0.30	0.18	0.20	0.12	1.43	1.42	1.42	1.23	1.22	1.22
AhPan	0.52		0.32	0.33	0.30	1.43	1.42	1.42	1.26	1.25	1.25
AhMon	0.31	0.63		0.21	0.19	1.41	1.39	1.40	1.19	1.19	1.18
AhKis	0.49	0.81	0.60		0.19	1.41	1.39	1.40	1.24	1.24	1.23
AhKor	0.16	0.52	0.34	0.49		1.41	1.40	1.41	1.24	1.23	1.23
AbTri	8.58	8.70	8.46	8.58	8.46		0.21	0.10	1.26	1.28	1.26
AbKai	8.62	8.74	8.50	8.62	8.50	0.60		0.17	1.29	1.31	1.29
AbAit	8.46	8.58	8.34	8.46	8.34	0.10	0.49		1.26	1.28	1.26
AbKor	6.94	7.30	6.83	7.26	7.06	7.63	8.12	7.63		0.25	0.17
AbKal	6.71	7.07	6.60	7.11	6.83	7.52	8.00	7.52	0.63		0.18
AbKip	6.60	6.95	6.48	6.99	6.71	7.28	7.76	7.28	0.31	0.31	

<b>Table 7.</b> Genetic distances (%) (below the diagonal) and standard errors (%) (above the diagonal) calculated for the control
region between the populations examined, considering the Kimura 2 parameter substitution model

	AhTin	AhPan	AhMon	AhKis	AhKor	AbTri	AbKai	AbAit	AbKor	AbKal	AbKip
AhTin		0.42	0.57	0.43	0.35	2.58	2.61	2.59	1.85	1.86	1.82
AhPan	1.28		0.42	0.44	0.39	2.52	2.55	2.54	1.81	1.82	1.79
AhMon	1.63	1.39		0.37	0.33	2.50	2.55	2.51	1.80	1.83	1.78
AhKis	1.49	1.68	1.32		0.41	2.51	2.55	2.53	1.78	1.81	1.77
AhKor	1.06	1.39	1.13	1.56		2.54	2.57	2.55	1.81	1.83	1.79
AbTri	24.34	23.86	23.48	23.89	24.08		0.23	0.15	1.80	1.80	1.82
AbKai	24.83	24.34	24.15	24.44	24.63	0.54		0.22	1.83	1.82	1.85
AbAit	24.53	24.05	23.69	24.10	24.29	0.39	0.54		1.82	1.82	1.83
AbKor	16.31	16.05	15.54	15.88	16.06	14.67	15.10	14.88		0.40	0.37
AbKal	16.40	16.14	15.89	16.15	16.32	14.41	14.75	14.58	1.51		0.46
AbKip	16.23	15.97	15.45	15.80	15.97	14.84	15.27	15.05	1.47	1.68	

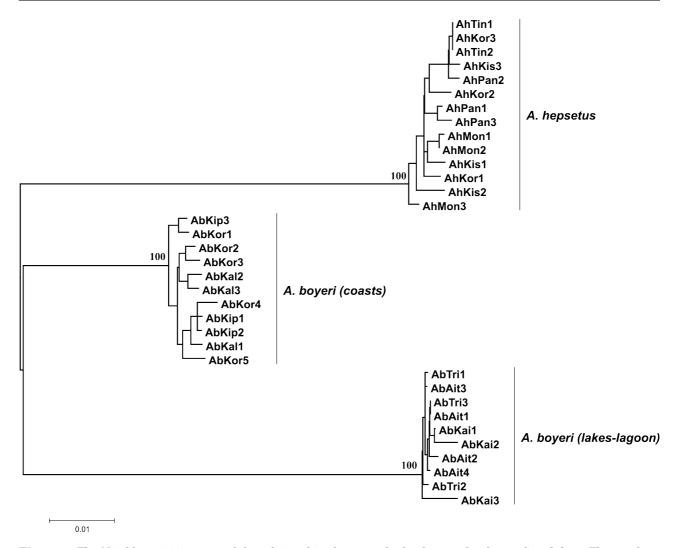
**Table 8.** Genetic distances (%) (below the diagonal) and standard errors (%) (above the diagonal) calculated for the combined data between the populations examined, considering the Kimura 2 parameter substitution model

	AhTin	AhPan	AhMon	AhKis	AhKor	AbTri	AbKai	AbAit	AbKor	AbKal	AbKip
AhTin		0.19	0.21	0.18	0.16	0.94	0.94	0.95	0.88	0.88	0.88
AhPan	0.64		0.19	0.21	0.19	0.93	0.93	0.93	0.88	0.89	0.88
AhMon	0.69	0.74		0.16	0.13	0.91	0.91	0.91	0.85	0.87	0.85
AhKis	0.72	0.91	0.71		0.17	0.91	0.91	0.91	0.86	0.88	0.86
AhKor	0.52	0.73	0.52	0.76		0.92	0.93	0.93	0.87	0.89	0.87
AbTri	13.00	12.91	12.73	12.86	12.89		0.11	0.06	0.78	0.77	0.77
AbKai	13.28	13.19	13.06	13.16	13.18	0.48		0.11	0.80	0.79	0.79
AbAit	13.03	12.94	12.77	12.90	12.92	0.17	0.45		0.79	0.78	0.77
AbKor	9.41	9.45	9.16	9.40	9.37	9.03	9.44	9.10		0.16	0.14
AbKal	9.41	9.44	9.24	9.48	9.44	8.88	9.26	8.94	0.79		0.17
AbKip	9.29	9.32	9.03	9.30	9.25	8.94	9.35	9.01	0.67	0.74	

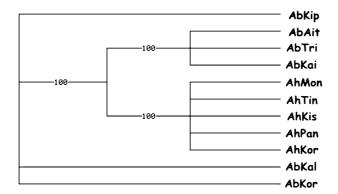
#### DISCUSSION

In the present study, we explored the genetic divergence and the phylogenetic relationships among different A. boyeri and A. hepsetus populations using mtDNA sequence analysis. The existence of three independent evolutionary units has been demonstrated by our data. The first one consists from A. hepsetus populations, whereas the A. boyeri populations are split into two separate genetic lineages, the first one containing the populations living in the sea and the second those living in the lakes and the lagoon. Genetic differentiation was found to be significantly high among those groups with the values observed between marine A. boyeri and A. boyeri living in the lagoon and the lakes being in the same order of magnitude as those estimated among the other two well defined species A. hepsetus and coastal A. boyeri; in any case, these values were found to be analogous to those existing among different species of the same genus (Karaiskou et al., 2003; Infante, Catanese & Manchando, 2004).

The clear genetic differentiation detected among the three groups mentioned above, is also highlighted by all phylogenetic trees constructed in this study. All trees drawn, either by NJ, MP or ML analysis for each of the three mtDNA segments examined, as well as for the combined data set, exhibited almost identical topology. All these methods resolved the same three distinct and deep divergent groups, a situation well supported from the high bootstrap values, reaching up to 100% in the majority of the cases. It must be noted that, our previous study (Klossa-Kilia et al., 2002), where the diversity of A. boyeri populations in Greece was tested using RFLP analysis, resulted in the conclusion that there are two different sibling species or at least subspecies or semispecies colonizing 'lagoon' and 'marine' habitats, respectively. Now it is clearer that the deep sequence divergence revealed in the present study, among the three groups tested, in combination with the fixed haplotypic differences found, suggests the occurrence of a different taxon in the A. boyeri species complex.



**Figure 2.** The Neighbour-joining tree of the relationships between the haplotypes for the combined data. The numbers at the tree nodes indicate the bootstrap values from 10 000 pseudoreplicates.



**Figure 3.** The maximum parsimony tree of the relationships between the populations for the combined data. The numbers at the tree nodes indicate the bootstrap values from 10 000 pseudoreplicates.

Several studies have recently shown that different fish ecotypes have experienced rapid genetic divergence and are, in many cases, on the brink of speciation (Taylor et al., 1997; Market et al., 1999; Lee, 2000; Staton, Foltz & Felder, 2000; Perdices et al., 2001). Significant genetic divergence has also been observed between estuarine and nearby marine populations in the silverside fish *Odontesthes argentinensis*. In this case, the authors assume that the colonization of estuarine habitats has promoted rapid adaptive divergence and reproductive isolation in estuarine populations, which are being considered as incipient ecological species (Beheregaray & Levy, 2000; Beheregaray & Sunnucks, 2001; Beheregaray, Sunnucks & Briscoe, 2002).

A growing body of evidence exists to suggest that estuarine life may promote speciation (Bilton, Paula &

Bishop, 2002). Thus, brackish water ecosystems are often exposed to a wide variation of environmental parameters such as temperature and salinity, which may cause strong selective pressures on organisms (Congiu et al., 2002). There is also evidence, pointing in the same direction, that estuarine environments tend to restrict gene flow imposing distinct selective regimes, which contributes to generate physiologically adapted populations, increasing the potential for in situ speciation in complete or partial isolation (Bilton et al., 2002). According to the same authors, the resulting taxa may be quite divergent from their marine counterparts and represent sibling or cryptic species of truly estuarine origin rather than simply estuarine populations of marine eurytopes. Furthermore, it appears that the phenomenon of specialized estuarine species does exist in a wide range of taxa, which, rather than being populated wholly by mobile marine eurytopes, often harbour endemic species whose recognition may have hampered by the morphological similarity to their fully marine siblings (Bilton et al., 2002).

By taking into consideration all of the above, we conclude that A. boyeri populations living in the lagoon as well as in the lakes are in essence allopatric compared with those of coastal A. boyeri ones. A similar situation has been recognized by many authors (Richardson, Baverstock & Adams, 1986; Allibone et al., 1996; McGlashan & Hughes, 2000) and it appears to make the determination of specific status under the biological species concept of Mayr (1942) rather difficult because there is no available evidence regarding the potential to inbreed in sympatry. Although the specific genetic status of the divergent A. boyeri 'lagoon' populations is difficult to judge definitively, the situation described, in combination with analogous ones in the literature, can be considered as generating difficulties regarding the biological species concept. However, apart from this theoretical discussion and taking into consideration our clear data, according to which the high genetic distances observed among marine and 'lagoon' A. boyeri populations are of the same order of magnitude as those found among the two well defined species A. hepsetus and coastal A. boyeri, as well as the finding that no haplotypes of marine origin were found among lagoon/lake A. boyeri populations, we tend to believe the existence of different species (cryptic or sibling) occupying freshwater environments rather than strongly differentiated populations of the same species. Similar data for A. boyeri populations were also provided by Trabelsi et al. (2002b).

A great amount of genetic divergence existing among *A. boyeri* populations originating from different (distant) lagoons has been reported by Trabelsi *et al.* (2004) and Astolfi *et al.* (2005). This was not confirmed in our case because the same haplotypes

appeared to be present in both lakes that were studied. This situation might be due to recent restocking activities, although there are no data available concerning this. In any case, more studies are needed to elucidate this finding by using more individuals, as well as more populations, from different lakes and lagoons in Greece.

Molecular genetic studies have revealed 'cryptic' taxa existing in several organisms, which exhibit considerable genetic divergence from their closest relatives and show little or no morphological differentiation (Jaarola & Searle, 2004). The existence of cryptic biological diversity has a wide range of implications for both ecological and evolutionary studies as well as for the preservation of biodiversity. Thus, the identification in our study of two discrete evolutionary lineages of *A. boyeri* at the intraspecific level has important implications for the conservation of this species because, from the management point of view, estuarine and marine populations should be treated as individual stocks when formulating fisheries policies (Beheregaray & Sunnuck, 2001).

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